

# Phenotypic and Genotypic Variability among Three *Bacillus megatherium* Isolates. 2- Molecular detection of Orthocide Fungicide Biodegradation

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**Abstract:** Three *B. megatherium* isolates (BM1, BM2 and BM3) showed a great ability to degrade the fungicidal orthocide 75 (95% Captan) in vitro and in vivo. They differed in their ability to orthocide degradation in vitro with 14, 10.8 and 9.0% for BM1, BM2 and BM3 isolates respectively in the course 48 hr incubation. In addition they differed in their ability to orthocide degradation in soil in the course 30 days after the fungicide treatment. Residual fungicide determined by GLC after 10, 20 and 30 days. It is clear that *B. megatherium* isolate B1 was the most active isolate degrading fungicide. The percentage of orthocide residual were 16, 20, 35 (non-sterilized soil) 43, 46 and 47% (sterilized soil) at 30 days for BM1, BM2 and BM3 isolates respectively. On the other hand, the effect of orthocide on *B. megatherium* growth was little (non-observed). *B. megatherium* isolates were varied in dehalogenase activity the data showed the level of enzyme activity was found to be considerably higher in BM1 isolate followed by BM2 and BM3. Dehalogenase gene was successfully detected in total DNA genome of three isolates by polymerase chain reaction (PCR), where as showed differences among isolates related to number, density and size of isozymes (bands). In addition base pairs nucleotides (gene) were one band (478 bp) BM1, 3 bands (1059, 560, 478 bp) BM2 and 3 bands (1300, 800, 478 bp) BM3, isolates relative mobility and density. The similarity of dehalogenase activity 20% among three *B. megatherium* isolates. [Journal of American Science 2010;6(8):116-121]. (ISSN: 1545-1003).

**Key words:** Fungicide biodegradation, Dehalogenase PCR; *B. megatherium*, GLC.

## 1. Introduction

The biodegradational behaviour of pesticides on and in various edible plants was studied by many authors. However, this type of study showed be carried out under the local conditions of every country. They concluded that, the current levels of the pesticides in surface water and plants at a not constitute on acute toxicity hazard to man on a short-term basis, but as a result of its accumulation in man body it causes hazard of large term exposure. On the other hand, some researches were published about the methods of removing pesticides from soil and water by activated carbon (Hegazi et al, 1990), biodegradation by microorganisms (Yakato et al, 1987, Dahrog et al., 2006, Amer, 2008 and Azhar, 2009).

So far haloalkane dehalogenases are the only enzymes known to be capable of direct hydrolytic dehalogenation of chlorinated and brominated hydrocarbons without the requirement for enzymes or oxygen. The enzyme of *Xanthomonas auroriphicus* G110's constitutively expressed 2 to 3%, of the soluble cellular protein (Keuning et al, 1985) up to 30% (Janssen et al, 1989). It has a remarkably broad substrate range which includes terminally halogenated alkanes with chain lengths up to 4

carbons for chlorinated and up to at least to carbons for brominated alkanes. Other haloalkane dehalogenases of bread substrate range have been found in gram positive haloalkane, utilizing bacteria (Janssen, et al, 1989).

## 2. Material and Methods

Microbial cultures:

Three *B. megatherium* isolates (namely BM1, BM2, BM3) from different soils were grown on nutrient broth medium. These isolates were obtained from microbiology Lab. Botany Dept., Fac. Sci., Benha Univ. (Nahid, 2009).

Fungicide:

Orthocide 75 (95% captan): N-Trichloromethyl-mercapto-4-Cyclohexan-1,2-dicarboximide, supplied by Kumia Chemical Industries Co. Ltd., Japan. Emulsifiable of the active ingredient 40% concentration of this orthocide was used in this investigation.

Detection of orthocide degradation:

In Vitro: Orthocide clearing zone: Assay medium contained (g/L): 6.0 NaCl, 1.0 (NH<sub>4</sub>)SO<sub>4</sub>; 0.5 KH<sub>2</sub>PO<sub>4</sub>; 0.1 CaCl<sub>2</sub>, 1.0 yeast extract and d H<sub>2</sub>O

up to one liter according to Vincent (1970) and supplemented with 0.5% (V/V) of emulsified orthocide (50%). This assay was carried out to determine the most potent degradation on the basis of mean diameters of clean zones (mm) by the testing *B. megatherium* isolates. After 24 hr petri-dishes were inspected to estimate the efficacy of orthocide degradation by millimeter ruler.

In vivo: Potted experimentally :

The soil taken from the surface layer (20 cm depth) of clay loam soil. Fourty pots (250 ml capacity) were filled with soil (about 200 pot-1). Twenty pots were sterilized and an other twenty pots were left without sterilization. The soil moisture content in all pots was raised up to 60% of WHC at 15 lbs 21 hr during 2 days. All pots were supplemented with 5000 ppm of orthocide and 100 ml broth media and mixed. Twenty ml of each broth medium contained ~ 5x10<sup>8</sup> cfu of three *B. megatherium* isolates used for inoculation of the soil and soil without inoculation as control. Soil samples were taken periodically after, 10, 20 and 30 days for the determination of residual orthocide using Gas liquid chromatography (GLC) according to Vogeler (1968). As well as counted *B. megatherium* in sterilized and non-sterilized soil using plate count method. The soil samples were kept in freezer until analysis to determine the residual amount of orthocide.

Determination of dihalogenase activity and isozyme using polymerase chain reaction (PCR):

Enzyme activity: Cell cultures were harvested at the end of the experimental growth phase. The protein was extract from cell cultures as described by Bradford (1976). The enzyme activity was assayed according to Janssen et al (1987).

Isolation of genomic DNA:

Bacterial cells were cultured in broth medium and incubated at 28°C with shaking at 200 rpm for 24 hr. and harvested by centrifugation. The DNA was extracted using CTAB method as described by Owen and Borman (1987).

Amplification of DNA:

The DNA was amplified by polymease chain reaction (PCR) in 100 µl react mixture containing: 20 µl template DNA (25 mg), 0.2 µl Taq DNA polymerase (unit); 3.0 µl DNTP, (25 mM of each dATP, dCTP, dTTP, dGTP), 3.0 µl MgCl<sub>2</sub> (25 mM); 30 µl PCR buffer (10X); 20 µl specific primer (dehalogenase encoding gen) Table (1) and 23.8 µl dH<sub>2</sub>O. The mixture was assembled on ice overlaid with 2 drop of mineral oil. The amplification was

carried out in DNA thermal cycler (MWG-Biko-TECH Primuse) programmed as follows: One cycle at 94°C for 2 min and then 30 cycles at 94°C for 30 sec., 55°C for 30 sec. and 72°C for 20 sec. On cycle at 72°C for 5 min then store in 4°C final mix of PCR.

**Table (1): Oligonucleotide sequences specific primers.**

Primer sets	5` Sequence 3`
TF	tgggcggattttggggct
TR	gtacgaatggccagcgtcc

Gel electrophoresis analysis: Agarose gel (1.5%) was prepared in TAE buffer. Total sample, 10 µl miniprep; 2 µl 6x loading dye and 6 µl dH<sub>2</sub>O of each isolate was loaded in each well. The electrophoresis was done of 65 V for 1 hr and then stained with ethidium promide solution (10 mg/ml) for about 10-15 min. DNA amplified was visualized on UV transilluminator and photographed.

Determination of orthocide residues:

1-Extraction and clean up:

One gram of each treated soil and untreated soil for each isolates were transferred into 250 ml separating funnel and added 60 ml at 15% methylene chloride/hexane (V/V). The funnel was stoppered and shaken vigorously for 2 min and vented the pressure during shaking. The two layers were separated. The aqueous layer was drawn off into separating funnel. This procedure was repeated using 30 ml of 15% methylene chloride/hexane (V/V). The aqueous layer was discarded and the solvent extracts in both separating funnels were combined and poured on top of anhydrous sodium sulphate column. Before the solution recedes the top of sodium sulphate layer, three 10 ml rinses of 15% methylene chloride/hexane were added. The filtrates were collected for clean up by using the US.EPA, 1988 procedures.

For clean up on a florisil column chromatography (300 mm long x 25 mm) internal diameter (i.d) with a small glass plug in the bottom was prepared by adding an activated florisil (130°C/overnight) in small portions, while tabbing until about 10 cm high. About one and a half cm layer at anhydrous sodium sulphate was added to the top carefully without mixing with florisil.

Mature solvents of 6% eiethyl ether-petroleum ether was added then 10 ml of 15% diethyl ether/petroleum ether and 50% diethyl/petroleum ether were used for elution.

The filtered elution mixture was evaporated using rotary evaporator 40°C for gas chromatographic determination.

## 2. Determination:

Separating of the residues was done on a borosilicate glass column 2 meters long, 3 mm internal diameter containing 3% oV-17 (phenyl methyl silicone) on chromosorb W, H.b. 100/120 mesh. The operating conditions were column oven temperature (150°C) carrier gas (Nitrogen) (1.2 kg/cm), Burner gas (Hydrogen) (1.0 kg/cm), and air (1.0 kg/cm). Phillips PU4410, computerized (Gas Chromatography with FPD) was used.

Soil free from Linuron was used to estimate rate of recovery by using the previous producers. The mean of the obtained recovery was 88%. All the obtained data for the residues of Linuron on treated soil were corrected by using such rate of recovery.

A series of concentration 10, 20, 30, 40, 50, 60, 70 mg of A falon analytical standard to 10 and toluene were prepared for obtaining the standard curve (Fig. 1). A suitable aliquot (5 µl) was injected from each concentration.

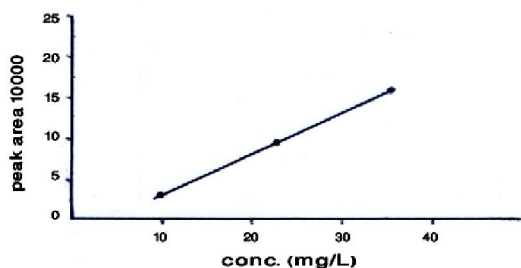


Fig. (1). Standard calibration curve orthocicide.

## 3. Results

*Bacillus megatherium* isolates under study showed a great ability to degrade fungicide orthocicide. They

differed in their ability to orthocicide degradation *in vitro*. Data concerning the amount of orthocicide degradation with 0.70; 0.54 and 0.44 gm with 14, 10.8, 0.9% for B1, B2 and B3 isolates respectively (Table, 2).

**Table (2).**The potent ability of *B. megatherium* isolates to orthocicide degradation *in vitro*.

B. megatherium isolates	Diameter clearing zone (cm)	Amount of orthocicide degradation	Percent of degradation (%)
BM1	3.5	0.70	14
BM2	2.7	0.54	10.8
BM3	2.2	0.44	9.0

Orthocicide amount in plate 5%.

The biodegradation of orthocicide *in vitro* in sterilized and non-sterilized soil inoculated with *B. megatherium* isolates (BM1, BM2, BM3) showed that the amount of orthocicide was decreased gradually through 30 days in sterilized and non sterilized soils (Table 3). The residual concentration of orthocicide were 35, 40 and 47 ppm in sterilized soil for inoculated, B1, B2 and B3 respectively (Table 3). While the orthocicide concentration was completely disappeared in non-sterilized inoculated soils were 16, 20 and 35 ppm after 30 days post inoculation (Table 3).

On the other hand, the total count of *B. megatherium* isolates were decreased in the first days then increased gradually under sterilized and non-sterilized soils (Table 3). The number of *B. megatherium* isolates were  $5 \times 10^7$ ,  $4 \times 10^7$  and  $3 \times 10^7$  cfu at 30 day post inoculation with BM1, BM2 and BM3 respectively.

Table (3): Biodegradation of orthocicide fungicide in soil inoculated with *B. megatherium* isolates.

Isolates	Non-sterilized soil			Sterilized soil		
	Orthocicide (ppm)	% of orthocicide residue	Total count	Orthocicide (ppm)	% of orthocicide residue	Total count
BM1 10th	4250	89	$1 \times 10^6$	4525	91	$2 \times 10^6$
20th	2750	55	$2 \times 10^6$	3375	68	$3 \times 10^6$
30th	775	16	$5 \times 10^7$	1750	35	$5 \times 10^8$
BM1 10th	4325	88	$2 \times 10^6$	4500	90	$2 \times 10^6$
20th	2975	60	$3 \times 10^7$	3605	82	$3 \times 10^7$
30th	1050	20	$4 \times 10^7$	2000	40	$3 \times 10^8$
BM1 10th	4350	87	$1 \times 10^6$	4625	93	$2 \times 10^6$
20th	3.50	61	$2 \times 10^7$	3627	60	$4 \times 10^7$
30th	1520	35	$3 \times 10^7$	2375	47	$4 \times 10^8$

\* After one hour from addition.

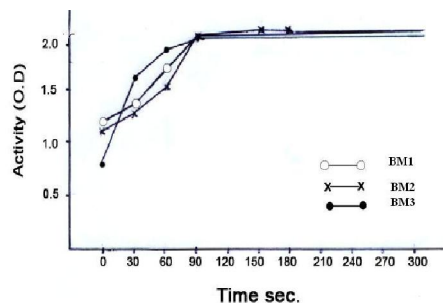
\*\* % of orthocicide residual related to control (without inoculated).30 days post inoculation.

Dehalogenase enzyme analysis:

Dehalogenase activity was varied among three *B. megatherium* isolates. The results in Table (4) reveal the enzyme activity of three isolates. It was found higher in isolate BM1 240 followed by BM2 180 and BM3, 120 sec. (Fig. 2).

**Table (4). Rate of dehalogenase activity among *B. megatherium* isolates.**

Isolates	Dehalogenase activity		
	BM1 isolate	BM2 isolate	BM3 isolate
0 time	1.25	1.05	0.70
60 sec.	1.75	1.50	1.10
120 sec.	2.20	1.95	1.50
180 sec.	2.65	2.10	1.50
240 sec.	3.25	2.10	1.50
300 sec.	2.25	2.10	1.50
360 sec.	2.25	2.10	1.50



**Fig. (2). Showing the rate of dehalogenase activity among three *B. megatherium* isolates.**

Dehalogenase isozymes :

The quantity of DNA purified extracted from 3 *B. megatherium* isolates were confirmed by UV spectrophotometer. The DNA concentration were about 30 µg/0.5 g cells. The purity of DNA was measured on 260/280 absorbance 1:65, 1:75 and 1:50 for 3 isolates respectively, these results indicating high yield and DNA purity:

The dehalogenase gene of 3 isolates were amplified of DNA, of 3 isolates using specific primers, taq polymerase and PCR reaction mixture using one set PCR method.

PCR amplification of dehalogenase gene appeared variation among 3 *B. megatherium* isolates, i.e. number, density and size of PCR products. The results in Table 5 and Fig. 3 showed total 7 bands of dehalogenase isozymes in all isolates as well as one bands of dhl clone (505 bp) of dehalogenase. *B. megatherium* isolate B1 has one isozyme band with 478 bp and both isolates BM2, BM3 has three isozyme bands with (1059; 560; 478) and (1300, 800 and 478 bp). On the other hand, these isozyme bands were differed in density (percent of fraction). The variability among 3 isolates, one monomorphic (common) amplified band (590 bp with 14% percentage and two polymorphic (specific) amplified bands (1540 and 725 bp) with 29% percentage. The different among three *B. megatherium* isolate related to genetic variability.

**Table (5). Genetic variability (DNA markers) of dehalogenase gene among, three *B. megatherium* isolates using PCR.**

Rf	Dhl %	Clone MW	<i>B. megatherium</i> isolates						Polymorphism
			BM1		BM2		BM3		
			%	MW	%	MW	%	MW	
0.28	-	-	-	-	-	-	54.7	1300	Unique
0.32	-	-	-	-	15.5	1059	-	-	Unique
0.34	-	-	-	-	-	-	18.3	800	Unique
0.39	100	505	-	-	78.7	560	-	-	Unique
0.42	-	-	100	478	5.8	478	27.0	478	Mono-morphic

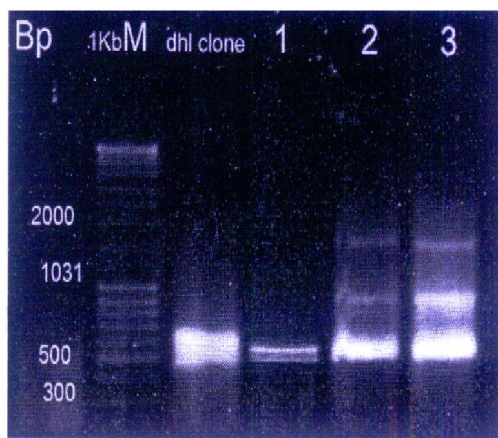
Rf : Relative mobility.

% : Percent of amplified band (DNA fraction)

MW: Size of expect band (bp).

Unique : Genetic marker

Monomorphic or common amplified band.



**Fig. 3. Agarose gel electrophoresis stained with ethidium bromide showing the PCR products of amplified dehalogenase gene of three isolates using specific primers, Lane M: DNA molecular weight marker, Lanes gene clone, BM1, BM2 and BM3 B. megatherium isolates.**

#### 4. Discussion

In the recent years, great quantities of fungicides are consumed annually for the control of plant pathogenic fungi during our chemical warfare against a multitude of noxious organisms in the soil. It is necessary to avoid the injury of these fungicides together with their various carriers diluents and solvents on the various beneficial soil microorganisms and their biological activities contributing to soil fertility.

*B. megatherium* isolates showed great ability to degrade orthocide fungicide in vitro and in vivo which assimilation as a source of carbon. The analysis of orthocide residual by GLC at 10, 20 and 30 days post-inoculation showed the decrease gradually of orthocide at 20 days in sterilized and non-sterilized soils.

The results indicate that, the biodegradation of orthocide residues in inoculated sterilized soil was high rate (16, 20, 25) followed by non-sterilized soil (43, 46, 47%) with B1, B2, B3 isolates respectively. This difference due to microbial flora in soil may be antagonistic with *B. megatherium* as well as it had the role in biodegradation of orthocide. In addition, this may be attributed to the concentration of soil (from drift occurrence of this fungicide).

Biodegradation of orthocide was detected in vitro with bacterial isolates but in different values

(0.70, 0.54 and 0.44 ppm) of B1, B2 and B3 isolates respectively.

This obtained previous results are in agreement with those obtained by Hashish et al. (1990); Dahrog, et al. (2006), Amer, (2008) and Azhar (2009) studied the ability *Pseudomonas*, *Bacillus*, *Rhizobium* and *Streptomyces* to degrade the Linuron, Diuron, Ktozin and dichlofuanid. The actual degrade the fungicides and herbicides by microorganisms is caused by the release of enzymes that breakdown of them. The role of fungicide, herbicide and pesticide on bacterial population in treated soil was intensively studied the reported by Berger (1998), bacteria capable of significant biotransformation and reduced phenylurea concentrations in liquid culture.

Shin and Cheney (1989) and Amer, (2008), conducted a trial to determine the effect of Linuron, Simazine, alcohol and nonselective parquet on *Bradyrhizobium japonicum* and *Rh. Leguminosarum* bacteria. The alcohol and Linuron were decreased significantly by 27.4 and 57.8% respectively, while little effect was observed in simazine and marked reduction of survivals observed in 200 ppm of parquet.

The degradation of orthocide by *B. megatherium* isolates proceeds through the concerted enzyme actions that are specific for halogenated compound and enzymes that are involved in the metabolism of natural compounds. The former are the dehalogenase that catalyze hydrolysis of orthocide. The enzymes show a broad substrate range and one only produced by isolates that utilize halogenated substrates and thus can be considered enzymes that are acquired by this specific strain of *B. megatherium* during genetic adaptation to degradation of chlorinated hydrocarbons. This paper describe present a further are lysis of the data alkane dehalogenase encoding gene *dhIA*. The absence of dehalogenase activities in natural of *B. megatherium* allowed the identification of clones containing the dehalogenase gene. Isolation of genes involved in methanol-dehalogenase and chloroacetaldehyde dehydrogenase activity was possible by screening for complementation of mutants lacking the dehalogenase activities. In this way; harboring genes were identified and the genes were localized on different DNA segments.

The efficient expression of the *dhIA* gene in other gram-positive bacteria is not surprising in view of the fact that two regions with the consensus *E. coli* promoter sequence were present. Copy number probably also plays a role, since expression levels were higher in *Xanthobacter autotrophicus* G10 (p 120) than in the wild type

isolates G110. The *E. coli* consensus promoter sequence is known to stimulate transcription in *B. megatherium*, (Dpyle et al, 1984 and Jeenes, et al. 1986) and our data suggest that it might also do so in *Pseudomonas* sp. In order to determine which of these sequences is the actual cause of the higher expression and whether the promoter can be used for expression of others genes in *Pseudomonas* spp., it will be necessary to identify the transcription short site of the gene and to study expression of different genes linked to the promoter regions.

## 5. References

- Amer, M.M. (2008). Monitoring of variation among faba bean *Rhizobium* isolates 2-Biodegradation of Herbicide 3 (3,4-ducgkiriogebatk)-1-methoxy-1methylurea. *J. Applied Science Rearch* 4 (6): 1001-1017.
- Azhar, A. El-Sayed (2009). Molecular detection of Biodegradation of Diuron herbicide in Relation to *Pseudomonas* spp.
- Berger, B.M. (1998). Parameters influencing biotransformation rates of phenylurea herbicides by soil microorganisms. *Pesticide Biochem. Physiol.*, 60 (2) 71-82.
- Bradford, M.M. (1976). A rapid and sensitive method for the quantification of microgram quantities of protein utilizing the principle of protein dye binding. *Analyt. Biochem.*, 72: 248-254.
- Dahrog, S.M.A.; W.M.A. El-Sayed; A.M. Abd-Allah and H.S.I. El-Masselati (2006). Degradation of linuron and metribuzi herbicides in relation to microbial population of *Bacillus* and *Pseudomonas* spp. *J. of Environmental Science*, Vol. (13), No. 1: 131-149.
- Dpyle, J.L.; D.N. Nunn and M.E. Lidstrom (1984). Molecular cloning of malyl coenzyme A lyase gene from *Pseudomonas* sp. strain AMI-a facultative methyltroph. *J. Bacteriol.* 160: 718-723.
- Hashish, R.M.; M.A. Azazy and A.N. Ibrahim (1990). Biodegradation of fungicide Kitazin-p by soil microorganisms in relation to its residues in rice grains. *Annals Agric. Sci., Fac. Agric., Ain Shams Univ., Egypt* (2): 653-665.
- Hegazi, M.E.; A.G. El-Sisi; M.M. Abou-Zatim and M.M. Diab (1990). Persistence of Dursban and Efficiency of some suggested methods to remove it from water. *Annals Agric., Fac. Agric., Ain Shams Univ.* 35 (2): 1057-1065.
- Jassen, D.B.; C. Kuning and B. Withold (1987). Involvement of a quinoprotein alcohol dehydrogenase and an NAD-development aldehyde ehalogenase in 2-chloroethanol metabolism in *Xanthobacter autotrophicus* G110. *J. Gen. Microbial. Microbial.* 133: 85-92.
- Janssen, D.B.; F. Pries; J. Ploeg; B. Razemier; P. Terpstra and B. Sitholt (1989). Cloning of 1,2-dichloroethane degradation genes of *Xanthobacter autotrophicus* G110 and expression and sequencing of *dh1A* gene. *J. Bacteriol.* P. 6741-6799.
- Jeenes, D.J.; L. Soldatri; H. Baur; J.M. Walson; A. Mercenler; C. Relammann; T. Leisinger and D. Haus (1986). Expression of biosynthetic genes from *P. aeruginosa* and *E. coli* in the heterologous host. *Mol. Gen. Genet.* 203: 241-429.
- Keuning, S.; D.B. Janssen and B. Witholt (1985). Purification and characterization of hydrolytic halokalane dehalogenas from *Xanthobacter autotrophicus* GJ10. *J. Bacteriol.* 163: 635-663.
- Nahed, A. Aiat (2009). Phenotypic and genotypic variability among three *Bacillus megatherium* isolates. I. In vitro evaluation of tri-calcium phosphate solubilizing potential and growth pattern (in press).
- Owen, R.J. and P. Borman (1987). A rapid biochemical method for purifying high molecular weight bacterial high molecular weight bacterial chromosomal DNA for restriction enzyme analysis. *Nucleic acids. Res.* 15: 3631.
- Shin, J.Y. and M.A. Cheney (1989). Abiotic dealkylation and hydrolysis of atrazine by brinessite. *Environ. Toxicol., Chem.*, 24 (6): 1343-1360.
- Vincent, J.M. (1970). A manual for the practical study of the root nodule bacteria. International biological programme 7 Mary Lobone Rood London Ni Blackwell Scientific Publication. Oxford and Edinburgh, P. 75.
- Vogeler, K. (1968). Gas chromatography method for determining residues of Hisosan in rice. *Pflanzenschutz-Nachrichten Bayer*, 211 (317-32).
- Yakota, T.; T. Omori and T. Kodams (1987). Purification and properties of haloalkane dehalogenase from *Corybacterium* sp. strain ml 5.3. *J. Bacteriol.* 169: 4049-4054.

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