Determination of Natural Radioactivity in Drinking Water and Consequent Dose to Public

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Abstract: Tap potable water in Egypt is a necessity practice, rather than a choice. Radium as well as other heavy metals in drinking water can pose a health hazard to human. In this work, determination of natural radium isotopes (226 Ra and 228 Ra) activity concentrations (mBq/L) in potable water samples from various locations in Egypt were carried out. The effective dose (mSv/y) and the associated cancer risk to public were estimated. The activity concentrations of 226 Ra were found to be in the range of 0.5 ± 0.2 to 22.0 ± 1.3 mBq/L with an average 3.6 ± 0.4 mBq/L. The activity concentration of 228 Ra were in the range of 41.6 ± 5.19 to 116.8 ± 14.6 mBq/l with an average 57.97 ± 9.49 mBq/l. The average estimated effective doses due to water consumption were 0.73 µSv/y for 226 Ra and 38.26 µSv/y for 228 Ra. The cancer risks due to water consumption during the life time (70 y) estimated was found to be in the range value of 7.89×10^{-5} .

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INTRODUCTION

Determination of radium isotopes in water has become a matter of interest in public health. Also determination of long-lived ²²⁶Ra and its progenies ²¹⁰Pb and ²¹⁰Po is of great interest in the field of radioprotection. These radionuclides preferentially accumulate in bones leading to rather long residence times in the human body. The radioactivity of the water intended for human consumption has been brought to public attention by recent international regulations such as EU council Directive 98/83/EC [1], Euratom [2,3] and WHO 1998 [4,5]. Upon ingestion of radium containing water, however, part of the radium is absorbed from the digestive tract and preferentially deposits in the bone. Radiation emitted by the radium isotopes and /or their progenies in the form of alpha or beta particles may then damage tissue and, over time, lead to the development of bone cancer.

The U.S. Environmental Protection Agency (USEPA) has established a Maximum Contaminant Level (MCL) for radium isotopes in public drinkingwater supplies because of health risks associated with its ingestion [6]. The MCL is 5 pCi/L for combined radium, which is defined as the sum of ²²⁶Ra and ²²⁸Ra. The contribution of drinking water to total human exposure is growing due to the presence of naturally occurring radionuclides in both uranium and thorium decay series. An increasing concern presently exists in several countries for the determination of the radioactivity levels in drinking water from various resources by different routine control measurements. Four naturally occurring radium isotopes are among the nuclides pertaining to these decay chains. Of these, ²²⁶Ra is an α-emitter with a long half-life of 1622 y and closely follows calcium metabolism in human body with eventual deposition in bones. This might lead to the buildup of ²²⁶Ra and its daughters, causing potential health implications and high degree of radio-toxicity due to long exposure hazards. Following the same behavior, ²²⁸Ra is a β-emitter with comparatively long half-life (5.77 y) with considerable significance in public health related studies [6]. Details of the dose assessment methodology are always implemented [7] and extensive estimates are referred to define recommended exposure levels for the population in several countries.

The metabolic behavior of radium on the human body is similar to calcium: around 20% of the consumed radium remains in the body, mainly in bones [8]. The behavior of radium in the human body is one of the reasons why radium has a high value of committed effective dose per unit intake [9]. Furthermore, the relative relevance of the radium isotopes regarding ingestion dose conversion factors has changed over time. While in 1992, the dose factor for ²²⁶Ra and ²²⁸Ra were similar, ²²⁸Ra value established after 1996 is twice that of ²²⁶Ra (6.9×10⁻⁷ and 2.8×10⁻⁷ SvBq⁻¹ respectively) and is similar to that of ²¹⁰Pb [9].

There is a growing interest in the development of improved methods for the detection of Ra isotopes in water. Because low levels of radium are ordinarily encountered in environmental water samples, the determination of radium initially requires one or more preliminary separation and/or pre-concentration steps, both to remove elements which may interfere with counting and also to get rid of large quantities of inactive substances. Methods based on gamma spectrometry offer good sensitivity but they are time consuming and require the use of large water volumes to attain detection limits of some mBq/L. Therefore, even a very low background is not sufficient to reach the required detection limits. Here, radiochemical separation techniques which separate the ²²⁶Ra and ²²⁸Ra from the matrix can improve the detection limits. The procedure followed in this study for ²²⁸Ra and ²²⁶Ra determination was a pre-concentration of radium isotopes (and ¹³³Ba as a yield tracer) via adsorption onto MnO₂ precipitate, extraction of radium isotopes with diphonix resin followed by a second extraction of ²²⁸Ac to measure by liquid scintillation counting, LSC and finally, ²²⁶Ra co-precipitation with Barium sulphate to measure by an Alpha detector [10].

This study concerned with the determination of activity concentrations of ²²⁶Ra and ²²⁸Ra in eighteen water samples collected from nine Egyptian cities. A conservative approach used to assess total effective dose and the cancer risk was estimated based on USEPA approach.

EXPERIMENTAL

Apparatus: All samples were counted by alpha spectrometry and liquid scintillation counting, LSC. The α -counting was performed on an alpha spectrometr Model 7401/7401VR, while β -counting was carried out by The LSC using Packard TRI CARB 3170 TR/SL, which used pulse shape discrimination (PSD) for α/β separation. Additionally the background is reduced by the presence of a guard of bismuth germinate (BGO), which surrounds the counting chamber.

Samples: Eighteen tape water samples were collected from nine Egyptian towns. The samples were acidified and transferred to the laboratory. The water samples (one liter volume) were evaporated to a volume of 100 ml, then, it transferred into a Marinillie beaker for gamma spectrometric counting followed by the radiochemical separation of the investigated isotopes.

Procedures: Radiochemical separation of ²²⁶Ra and ²²⁶Ra isotopes was performed on all samples. MnO₂ co-precipitation at prepared using 0.015 g KMnO₄ and 0.5 M MnCl₂ the pH was adjusted between pH 8–9. MnO₂ used for the pre-concentration of radium isotopes and effectively separates some possible interfering radionuclides such as U isotopes and ⁹⁰Sr [10]. The Precipitate is dissolved in ~ 2ml concentrated HCl, which is then diluted to about 2M HCl by addition of 10 ml of water. An initial Diphonix resin has a strong affinity for actinides and lanthanides, and is used to decontaminate the sample of potential environmental interferences while allowing divalent

cations such as the radium isotopes and barium to pass through. The decontaminated load and rinse solution containing the Ra (Ba) fraction is collected in a clean glass beaker and held for at least 30 h to allow for ²²⁸Ac in-growth from the ²²⁸Ra in the sample. The equilibrium activity of ²²⁸Ac from ²²⁸Ra in the sample is then loaded onto a second Diphonix resin column. The load and rinse solutions from the 2nd column is collected in a clean beaker and is processed for the yield determination by measuring the ¹³³Ba gamma peak at 356 keV via gamma spectrometry. The ²²⁸Ac is then eluted from the second Diphonix column using (1 M 1-Hydroxyethane-1,1-diphosphonic acid) into a LSC vial and completed with UGAB liquid scintillation cocktail for measurement via LSC model TRI CARB 3170 TR/SI.

²²⁶Ra precipitated as Ra(Ba)SO₄ and determined by an alpha spectrometric model 7401/7401VR Alpha spectrometer system equipment with CANBERRA chamber of low background, high resolution PIPS detectors ion implanted detectors with 1200 mm² active area, with counting time not less than 80,000 seconds. The overall yield of this procedure is typically greater than 95%. Alpha-emitting radium isotopes (²²³Ra, ²²⁴Ra and ²²⁶Ra), now also free of radioactive interferences, may be analyzed by alpha spectrometry after microprecipitation with BaSO₄ from the final load and rinse solution of the 2nd Diphonix column.

For quality control, water reference materials IAEA-426 and IAEA-423 were analyzed during the same experiment run and the concentration of ²²⁶Ra and ²²⁸Ra were determined. The uncertainties of the results were evaluated considering counting statistics and calibration error only.

DOSE ESTIMATIONS

The Annual Effective dose estimations were calculated based on the radium isotopes activity concentration and the annual water consumption, using the following formula:

Annual Effective dose $(\mu Sv/y) = Activity$ Concentration $(Bq/L) \times Annual water consumption$ $(L/y) \times Dose coefficient factor (\mu Sv/Bq)$

Where:

The dose coefficient for 226 Ra and 228 Ra are 0.28µSvBq⁻¹ and 0.69µSvBq⁻¹ respectively correspond to a reference dose level (RDL) of 0.1mSv/year, annual water consumption assumed to be 730 liters/year, [11].

The Health effects subcommittee believes that the risks associated with all radium species should be combined so that the total risk is known. The assumed average ratio from the occurrence data is used to determine the concentration of each radium isotope that would meet an acceptable risk level [12].

calculated as follows:
Risk = MCL
$$\times$$
 RC \times TWI

Where:

Risk = Lifetime cancer risk corresponding to MCL (unit less)

MCL= Maximum contaminant level (Bq/L)

RC = Mortality risk coefficient for 226 Ra (7.17 × 10⁻⁹ Bq⁻¹), and for 228 Ra (2.0 × 10⁻⁸ Bq⁻¹)

TWI = Total water intake (2 L/d \times 365.4 d/y \times 70 y).

RESULTS AND DISCUSSION

The activity concentration of ²²⁶Ra and ²²⁸Ra in mBq/L in water samples are given in Table 1. The activity concentrations range from 0.5±0.19 to 22±1.3 mBq/L with average 3.6 mBq/L for ²²⁶Ra and from < DL to 116.8 \pm 14.6 mBq/L with an average 75.9 mBq/L for ²²⁸Ra. The mean activity concentrations of both ²²⁶Ra and ²²⁸Ra are within those in drinking water in several other countries such as united states [13,14,15,16], Pakistan [17,18], Finland [19,20], France [21,22], Germany [23,24,25], Italy [26], Poland [27,28], Romania [29,30], Switzerland [31], Spain[32] and U.K. [33] as shown in Table 3 and Fig. 1 and Fig. 2. Water types originating from the Eastern Nile Delta area are characterized by low 226Ra levels and relatively high ²²⁸Ra activity, presumably due to the muddy agricultural nature of that area, [35] which is subject to water from several surface resources for irrigation, from this point of view, the slight increase of

²²⁸Ra in few cities may be attributed to the same reason. The results showed that, in general, radium isotopes activity concentration in drinking water samples did not exceed the maximum contaminant level (5 pCi/L) recommended by USEPA [36]. The average committed effective doses µSv/y due to water consumption is given in Table 2. The highest values of the committed effective dose per year were 4.50 μ Sv/y for 226 Ra and 58.83 μ Sv/y for 228 Ra. The average committed effective doses per year were 0.73 μ Sv/y and 38.27 $\mu Sv/y$ for ^{226}Ra and ^{228}Ra respectively. The radium isotopes (and here especially ²²⁸Ra) are responsible for the major part of the annual effective dose from ingestion water. The corresponding total committed effective dose obtained was 38.99 µSv/y. So, regular consumption of radium through drinking water, even at activity concentrations few times over the maximum recommended value, will not lead to an effective dose higher than recommended 0.1 mSv.

Risks associated with ingestion of radium isotopes (226 Ra and 228 Ra) for the concentrations and exposure periods discussed above are given in Table 2. The USEPA established a range of 1×10^{-4} to 1×10^{-6} as an acceptable cancer incidence risk in the Notice of data availability for radionuclides in drinking water that was published on April 21, 2000. It is noted that the resulting risks are the same order of magnitude as the EPA's target ceiling risk (10-4); however, the actual value is different for each radionuclide.

Table 1. The Estimated ²²⁶Ra and ²²⁸Ra Activities in the Collected Water Samples:

Cities		Radioactivity level (mBq/L)		The mean of Ra	The mean of Radioactivity level (mBq/L)	
	No.	²²⁶ Ra	²²⁸ Ra	²²⁶ Ra	²²⁸ Ra	Recovery (%)
Cairo	a	1.66 ±0.34	< DL	1.66 ±0.34	< DL	83
cuito	b	1.66 ± 0.34	< DL			
El Mansoura	a	1.17 ± 0.27	< DL	1.17 ± 0.27	< DL	98
	b	1.17 ±0.27	< DL			
Qualuab	а	1.69±0.29	72.53±9.06	1.69±0.29	72.53±9.06	98
-	b	1.69±0.29	72.53±9.06			
Octobar	а	0.92 ± 0.07	< DL	0.92 ± 0.07	< DL	96
	b	0.92 ± 0.07	< DL			
Alex	a	1.79 ±0.96	$116.80{\pm}14.6$	1.79 ± 0.96	116.80±14.6	94
	b	1.79 ±0.96	$116.80{\pm}14.6$			
Tanta	а	0.53 ±0.17	< DL	0.53 ± 0.17	< DL	78.9
	b	0.53 ±0.17	< DL			
Baniswif	а	0.50 ±0.19	41.55±5.19	0.50 ± 0.19	41.55±5.19	94
	b	0.50 ±0.19	41.55±5.19			
Sinai	а	22.0±1.28	73.00±9.12	22.0±1.28	73.00±9.12	77
	b	22.0±1.28	73.00±9.12			
Siwa	а	1.79 ± 0.36	< DL	1.79 ± 0.36	< DL	83
	b	1.79 ± 0.36	< DL			

DL. Represent the detection limits for 228 Ra $\sim 20mBq/L$

Table 2. The Average Committed Effective Doses and Cancer Risk Associated Due to Consumption of Water:

Cities	Annual Effective Committed dose (µSv/a)		The Radiological Risk	
	²²⁶ Ra	²²⁸ Ra	²²⁶ Ra x 10 ⁻⁷	²²⁸ Ra x 10 ⁻⁵
Cairo	0.34	ND	6.09	ND
El Mansoura	0.24	ND	4.29	ND
Qualuab (Treatment factory)	0.35	36.50	6.20	7.42
Octobar	0.19	ND	3.38	ND
Alex	0.37	58.80	6.57	11.95
Tanta	0.11	ND	1.94	ND
Baniswif	0.10	20.93	1.83	4.25
Sinai	4.50	36.77	80.70	7.47
Siwa	0.37	ND	3.57	ND

ND. Not Determined

Table 3. Concentrations of ²²⁶Ra and ²²⁸Ra Isotopes in Drinking Water.

Region / country		(mBq/ L) in drinking	Ref.	
	water			
	Ra-226	Ra-228		
North America United States	0.4-1.8	0-0.5	Cothern, C.R. etal., 1983, Fisenne, I.M. et al., 1987,	
United States	0.4-1.8	0-0.5	Holtzman, R.B., 1964 and McCurdy, D.E. et al ., 1981[13,14,15,16]	
Asia				
Pakistan	0.2-120		NEPA, 1990, Dang, H.S. et al., 1990 [17,18]	
Europe				
Finland	10-49000	18-570	Asikainen, M.1982 and Salonen, L.1994[19,20]	
France	7-700		Descamps, B.et al., 1988 and Pellerin, P. et al., 1980[21,22]	
Germany	1-1800		Bundesministeriumfür Umwelt 1994, Gans, I. 1985 and Glöbel, B. et al., 1980 [23,24,25]	
Italy	0.2-1200		Sgorbati, G et al., 1997 [26]	
Poland	1.7-4.5		Pietrzak-Flis, Z. et al., 1997 and Pietrzak-Flis, Z. et al., 1997 [27,28]	
Romania	1.7-4.5		Botezatu, E., 1994 and RSRP., 1994 [29,30]	
Switzerland	0-1500	0-200	SFOPH, 1997 [31]	
Spain	0-1500		Soto, J. et al., 1988 [32]	
Ū.K.	0-180		Bradley, E.J. 1993 [33]	
Africa				
Egypt	0.5-22	41-117	This studies	
Reference value				
UNSCARE	0.5	0.5	UNSCARE [34]	

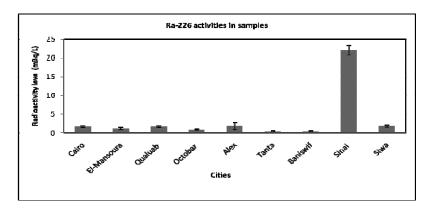


Fig. 1: The distribution of ²²⁶Ra activity concentration found in samples collected from different cities in Egypt

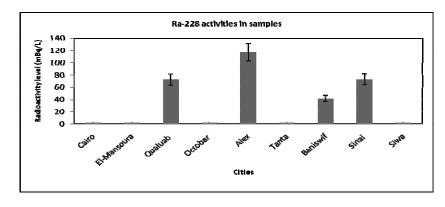


Fig. 2: The distribution of ²²⁸Ra activity concentration found in samples collected from different cities in Egypt

CONCLUSION

Concentration of ²²⁶Ra and ²²⁸Ra in potable water collected from nine Egyptian towns vary over a large range. The highest concentration of Radium isotopes are found in water types originating from wells such as Sinai and Siwa. The radiation doses estimated from natural radium in water are low compared with the average total dose of 2.4 mSv/y from external and internal radiation in EPA and WHO regulation. The average doses for the population are safely below the total indicative dose of 0.1 mSv/y imposed by the EU drinking water directive and WHO recommendation [5]. The risk assessment data show that the radionuclides under this investigation do not pose any significant health risk to the public.

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